

1 **Title:**

2 Estimation of the balance of heavy metals in Romanian agro-ecosystems and research  
3 directions

4 **Authors:**

5 Iordache Virgil<sup>a</sup>, Dumitru Mihai<sup>b</sup>

6 **Affiliations:**

7 <sup>a</sup>University of Bucharest, Department of Systems Ecology, Spl. Independentei 91-95,  
8 050089 Bucharest, Romania, virgil.iordache@bio.unibuc.ro

9 <sup>b</sup>Institute of Soil Science and Agrochemistry, Blv. Mărăști nr. 61, 71331 sector 1,  
10 Bucharest, Romania, mdumitru@icpa.ro

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12

13 **Abstract**

14 In order to characterize to current state of Romanian agro-systems with respect to heavy  
15 metals pollution and to identify mitigation measure, we performed an extensive review  
16 of the existing data, literature, and policies. The following areas were covered: the  
17 context of the problem of pollution with heavy metals in agro-systems, the background  
18 concentrations and non-agricultural inputs (heavy metals concentrations in soils,  
19 atmospheric deposition), agricultural practices and related inputs (mineral fertilizers,  
20 animal wastes, pesticides, irrigation water, flood-related inputs, livestock feeds), runoff  
21 and erosion, leaching, biological products, and animal excreta and waste water. Based  
22 on these data a heavy metal balance for Romania was performed. According to the  
23 existing data the major inputs of metals to agricultural lands are by natural fertilizers  
24 and atmospheric depositions, and the major outputs occur by dumped animal wastes and  
25 plant food products. Research priorities are discussed.

26 **Keywords:** Heavy metals; Agriculture; Balance; Romania.

27

28 **Content**

29 1 Inputs

30 1.1 Background concentrations and non-agricultural inputs

31 1.1.1 Heavy metals concentrations in soils

32 1.1.2 Atmospheric deposition

33 1.2 Agricultural practices and related inputs

34 1.2.1 Mineral fertilizers

35 1.2.2 Animal wastes

36 1.2.3 Industrial and urban sludge and composts

37 1.2.4 Pesticides

38 1.2.5 Irrigation water

39 1.2.6 Flood-related inputs

40 1.2.7 Livestock feeds

41 2 Outputs

42 2.1. Runoff and erosion

43 2.2 Leaching

44 2.3 Biological products

45 2.4 Animal excreta and waste water

46 3 Heavy metal balance for Romania

47 4 Research directions

48

49 **Introduction**

50 The problem of the pollution with heavy metals in agrosystems in Romania is framed by  
51 a context consisting of two elements.

52 1. the overall biogeochemical cycles of heavy metals in the natural and human

53 ecological systems of the country (the so called natural capital and socio-economic  
54 systems)

55 2. the overall deterioration of the agrosystems of the country

56

57 The first element of the context provides an idea of the relative importance of heavy  
58 metals pollution in agrosystems, as compared with the pollution of other types of  
59 ecological systems. It also makes explicit the external factors controlling the pollution  
60 of agrosystems (i.e. the factors originating in the connections between agrosystems and  
61 other natural and socio-economic systems). The second element of the context informs  
62 about the relative importance heavy metals pollution as compared with other types of  
63 deterioration occurring in the case of agrosystems.

64

65 *Element 1.* The national environmental strategy (MAPP, 1996) offers a synthesis of  
66 the current national environmental problems. One aspect pointed out is the heavy metals  
67 pollution of water and soil in different geographic areas of Romania. The main causes  
68 of pollution is related to several industrial platforms and mining areas, as well as wastes  
69 disposal. Most of the intensive pollution is usually limited to the dispersal area around  
70 these platforms and mining areas, and affects all types of ecological systems, including  
71 socio-economic systems (human settlements). Agrosystems represent an important type  
72 of ecological system in the intensively polluted area, but not the only one. The main  
73 mechanisms involved in the pollution of ecological systems: is dispersal by air, by  
74 rivers, and direct disposal. The dispersal by air is responsible for the pollution of most  
75 affected agrosystems, while dispersal by rivers is responsible for floodplain and  
76 downstream river contamination (including the relatively few agrosystems located in the  
77 floodplain). Monitoring reports show also that rivers biota (Lower Danube River fish)

78 have large concentrations of heavy metals far downstream the responsible industrial and  
79 mining sources (ICIM, 1999, INDD, 1999, Iordache et al., 2000), and sometimes even  
80 the abiotic compartments (Lacatusu et al, 1998). This downstream chronic pollution  
81 reflects all the sources from the Danube catchment, and by now there is no enough  
82 information allowing us to discriminate between the contribution of each country. Most  
83 of the studies concerning heavy metals pollution have adopted a sectoral approach (they  
84 have not envisaged the full complexity of the studied ecological systems). In particular,  
85 studies of soil pollution have dealt, with soil-plant systems, including litter, and only  
86 seldom animals (e.g. Lacatusu et al., 1997a). Information concerning other compartment  
87 have been extracted only from literature. Consequently, there is a lack of quantitative  
88 knowledge about the fluxes in which the different contaminated environmental factors  
89 (air, water, soil, biota) are involved, as well as about the cumulative and delayed effects  
90 in space and time. This sectoral approach is reflected also by the current monitoring  
91 system, with separate monitoring of the environmental factors by different responsible  
92 institutions, few communication between the involved actors, and separate strategies for  
93 improving the quality of the environmental factors (Varduca, 2002, Rauta et al., 1998,  
94 Cernatoni et al., 2002, Popescu and Vintila, 2002).

95

96 *Element 2.* An overall analyses of current priorities in soil protection is made in  
97 Dumitru et al. (2001). The analyses was based on an inventory performed on 12  
98 millions ha of the total agricultural area of 14856800 ha. Table 1 summarizes the  
99 surfaces of agricultural fields affected by different limiting factors of the productive  
100 capacity. The Zn deficiency registered over 1.5 millions ha significantly affects the  
101 maize production. On the other hand, very aggressive effects are due to the pollution  
102 with heavy metals (especially Cu, Pb, Zn, and Cd) and SO<sub>2</sub>, identified mainly in the

103 areas Baia Mare, Zlatna, Copsa Mica. Even if in the last years some industrial platforms  
104 have been closed (such as ROMFOSFOCHIM-Valea Calugareasca), and others have  
105 reduced their activities, soil pollution remains high also in areas such as Targu Mures,  
106 Turnu Magurele, Tulcea). In some of these areas there are documented toxicological  
107 (Dumitru, et al., 1996), and ecotoxicological (Vadineanu et al, 1991, quoted by Dumitru  
108 et al., 1996) effects of the heavy metals. Other approximately 700000 ha are covered  
109 with slight-moderately polluted soils (Lacatusu et al., 1992). In 1992 the area  
110 excessively polluted, 169242 ha, was smaller than in 2000 (200000 ha), and had as  
111 pollution causes those listed in table 2 (Vrinceanu et al., 1997). One can see that the  
112 excessive pollution due to agriculture plays minor role. However, agriculture may more  
113 important as a cause of slight-moderately polluted soils. According to Vrinceanu et al.  
114 (1997), soil pollution caused by agricultural technologies occurs either by strengthening  
115 of some negative natural physical and chemical characteristics of the soil (for example  
116 salt affection by salt waters or by the uncontrolled irrigation of soils with natural  
117 potential of salinization) or by the accumulation of the chemical substances used disease  
118 and pest control. Excessive fertilization has been another cause of soil pollution by  
119 technical means, especially the application of high rates of nitrogen fertilizers with high  
120 acidification potential and with high content of nitrate nitrogen. The deposition and  
121 application of the zootechnical wastes, of the sludge from city or industrial waste waters  
122 treatment plants is another source of soil contamination, when these operations are not  
123 carried out among the recommended limits for environmental protection (op.cit). It is  
124 worthy noting, as relevant for the input of metals by irrigation water, that the surface  
125 where irrigation infrastructure is available has continuously decrease from 1990  
126 (3215800 ha) to 2000 (400000). In 2000 only 204200 ha have been irrigated (\*\*\*,  
127 2001b). The restoration of soils and landscapes quality (including as priority the

128 rehabilitation of the soils heavily polluted with heavy metals) is evaluated as having a  
129 cost of ~25 billions Euro over 15-25 years (Dumitru et al., 2001).

130

131 In this context, it is not surprising that a detailed heavy metals balance at farm gate  
132 would not be possible to be computed at this stage (excepting for some experimental  
133 fields, with no relevance for the real farm systems), because of two basic reasons.  
134 Reason 1 relates to the sectoral approach dominating the research of metals  
135 biogeochemistry (which do not facilitate a focus on estimating fluxes connecting the  
136 agrosystems with other ecological systems), and is not the most effective in producing  
137 knowledge given a limited amount of financing for research. Reason 2 relates to the fact  
138 that there are other soil deterioration pathways as important as heavy metal pollution,  
139 which limited the local research funds dedicated to heavy metals pollution of  
140 agrosystems. Also, a balance at country level can be computed only with fragmentary  
141 data and by making assumptions which induce a high degree of risk to the decision  
142 made on the basis of these evaluations.

143

144 We will start now an analysis of each category of existing data.

145 *Table 1 about here*

146 *Table 2 about here*

## 147 **1 Inputs**

148 1.1 Background concentrations and non-agricultural inputs

149 1.1.1 Heavy metals concentrations in soils

150 The heavy metals concentrations in soil are comprehensively characterized, even before  
151 the start of the integrated soil monitoring system (Dumitru et al., 2000), by the mapping  
152 of the geogenic concentrations (e.g. Ianos, 1998, Lacatusu et al., 1993, 1994, 1997b,

153 2001, Mihailescu et al., 1986, 1988). These studies allowed the identification of areas  
154 with microelement deficiency, as well as of the areas with high natural background (the  
155 last ones are located mainly in mountainous areas of volcanic origin, with low relevance  
156 for the agrosystems). Table 3 present contents of potential polluting elements in top soil  
157 by main agricultural land uses and soil types.

158 *Table 3 about here*

### 159 1.1.2 Atmospheric deposition

160

161 Atmospheric deposition in Romania is monitored only in forest ecosystems, as part of  
162 the European specific monitoring network. However the fluxes analyzed in Romania  
163 include only S-SO<sub>4</sub>, N-NO<sub>3</sub>, N-NH<sub>4</sub>, Ca, Mg, K, and Na (Barbu 2004, Barbu et al.  
164 2004). For instances, the flux of S decreases from 16 kg/ha/year at low altitudes  
165 (~100m) to 8-10 kg/ha/year at altitudes above 1000m. The same pattern has been  
166 observed for Cl, N-NO<sub>3</sub>, and N-NH<sub>4</sub>. Average depositions of Ca, Mg, K, and N (in  
167 1998-2002) were 16 kg/ha/year, 3, 6.7, and 4.9, respectively, with generally higher  
168 values in the southern part of the country, and important year to year variations.

169 Quantitative data concerning heavy metals in air are available in what concerns  
170 emissions (Constantinescu, M., personal communication), but they cannot be indirectly  
171 used for assessing depositions due to the well known large scale atmospheric  
172 exchanges.

173

174 Indirect information concerning atmospheric deposition of heavy metals is provided by  
175 deposition models at European scale. For instance the Dobris report mention for  
176 Romania 0.5-1g/year of Cd over 80% of its surface, 0.2-0.5 g/ha/year over 17% of the  
177 surface and 1-2g/ha/year over 3% of the surface. The average values for Romania are, in



178 the mentioned report, more or less similar with the other countries in the area, such as  
179 Bulgaria, and slightly lower than Hungary. If we look at independent data characterizing  
180 Hungary (Bozo, 1993, Meszaros et al., 1993), we see that values for Cd range from  
181 1g/ha/year to 5.7 g/ha/year. This seems to confirm the predictions of the European scale  
182 models, and consequently we will use the average data characterizing Hungary adjusted  
183 by a fraction of 0.5 for a very rough assessment of deposition in Romanian soils.

184

185 On the other hand, it is known that the heavily polluted soils in the nearby of industrial  
186 platforms are polluted by air dispersal. However, direct quantification are not available  
187 in the literature.

188

## 189 1.2 Agricultural practices and related inputs

### 190 1.2.1 Mineral fertilizers

191 From a maximum of 1265299 t applied in 1986 (Dumitru et al., 2001), the amount of  
192 inorganic fertilizers have decreased in 2000 to 327537 t (\*\*\*, 2001a), of which 88258t  
193 are P fertilizers, with known capacity of inducing heavy metals accumulation in soils.

194 Mineral fertilizers have been applied in 2000 on 3724578 ha. The concentration of  
195 heavy metals in these fertilizers is unknown. In order to assess at least the magnitude  
196 order of the input by fertilizers, we will use an average of the values provided by studies  
197 performed in other European countries (Eckel and Dohler, 2001).

198

### 199 1.2.2 Animal wastes

200

201 In 2000 there were 21.4 cattle and 60.8 sheep and goats per ha of arable land, pastures  
202 and meadows, and there were 62.7 pigs per ha of arable land (\*\*\*, 2001a). Table 4

203 provides an assessment of total animal wastes produced in 2000, including the used  
204 assumptions. Rauta and Dumitru (1986), estimated that in 1982 the pig and poultry  
205 wastes included 403849 t of N, 98154 t of P, and 228258 t of K. Jinga et al. (1990)  
206 estimated the quantities of waste water associated to animal raising at that date as being  
207 53 mil m<sup>3</sup> from pigs, 15 mil. m<sup>3</sup> from cattle, 2 mil m<sup>3</sup> from poultry and 1.8 m<sup>3</sup> from  
208 sheep. We will characterize below the situation concerning the solid waste, the  
209 semiliquid wastes and the liquid wastes. Solid wastes refer to the sludge (which include  
210 both excreta and remnants of food), and in some cases to solid manure; liquid wastes  
211 refer to waste water (which include both diluted urine and substances physico-  
212 chemically partitioned from the sludge); semiliquid wastes refer to fresh pig excreta. All  
213 the available literature reporting metals concentrations dealt with solid and liquid wastes  
214 (tables 5 and 7).

215

216 A high number of experiments were performed with quantities of animal *sludge* ranging  
217 from 10t/ha to 100 t/ha (e.g. Popa, 1964, Jinga, 1978, Stepanescu et al., 1978-1980,  
218 Ionescu-Sisesti et al., 1980, 1981, Dumitru et al., 1982, \*\*\*, 1981-87, Dumitru et al,  
219 1993b). Up to 70 t/ha were found acceptable for pig sludge application over soil, but no  
220 less than 30 t/ha in order to obtain a significant increase in plant production. The  
221 quantity depended on soil type and culture. Nastea et al. (1983) recommended the  
222 application of 10 t of composted pig sludge / ha. In what concerns the cattle *solid*  
223 *manure*, 30t/ha of were found optimal (\*\*\*, 1981-87). In a syntheses paper, Jinga et al.  
224 (1990) recommend the application of 20-30 t of waste / ha, for obtaining a fertilizing  
225 effect with a duration of 1-2 years.

226

227 Jurubescu (1977) recommends an average of 50m<sup>3</sup> of *semiliquid wastes* / ha (20–75 m<sup>3</sup>

228 depending on soil and culture). The recommendations concerning the use of *waste*  
229 *water* are more variable, depending on the quality of the water and on the climatic area  
230 (Ionescu-Sisesti et al., 1979). Jinga (1971) reports a maximum of 7500 m<sup>3</sup>/ha in the  
231 studied case, and Jinga et al. (1984) recommend 2250-3875 m<sup>3</sup>/ha, with no more than  
232 500 m<sup>3</sup>/ha in one application. Frequently it is recommended a dilution of the waste  
233 water with normal irrigation water (e.g. Dumitru et al., 1982a, 500-800 m<sup>3</sup>/ha, 1:5  
234 diluted with normal water).

235

236 15812625 t of natural fertilizers were reported as applied over 647200 ha in Romania in  
237 2000 (\*\*\*, 2001a), which is below our estimation of animal waste production. No  
238 details separately concerning the application of solid, semiliquid and liquid animal  
239 waste are currently available at national level. For computing metals input by natural  
240 fertilizers we will use an average of the values mentioned in table 5.

241

### 242 1.2.3 Industrial and sewage sludge and composts

243 The total production of industrial sludge in 2000 was 2.5 milion t, of sewage sludge  
244 0.54 milion t, and of urban domestic wastes 8.15 milion t (\*\*\*, 2001a, Badea et al.,  
245 2003). In the 80-90 years a number of studies investigated the application of industrial  
246 and urban sludge (table 6, including concentrations of metals in these media). Dumitru  
247 et al. (1984) recommended the application of 50-60 t of industrial sludge / ha in the  
248 studied case, and Blaga et al. (1989) recommends the application of 60-80 t of industrial  
249 sludge /ha only after improving the technology in order to reduce the concentration of  
250 metals. Recently experiments with sewage sludge restarted (Borza and Radulescu,  
251 1998), in order to optimize the application solution for specific cities.

252

253

*Table 4, 5, 6, 7 about here*

254

Even if the application of sewage sludge is not yet regulated , ICPA (National Institute

255

for Soil Science and Agrochemistry), which is involved in advising and approving the

256

application of sewage sludge in case by case situations, currently uses the following

257

recommendations (Stefanescu and Dumitru, 1991):

258

- 6-10 t dm / ha every 3 years (30 t sewage sludge / ha with 65 % humidity every

259

three years) or

260

- 10-20 t dm / ha every 5 years (50 t sewage sludge /ha with 65 % humidity every 5

261

years).

262

263

Maximum acceptable metals concentrations in the applied sewage sludge are

264

(Stefanescu and Dumitru, 1991, in micrograms / gram dri weight): As 10, Cd 10, Co 10,

265

Cr 500, Cu 500, Mn 1200, Hg 10, Mo 20, Ni 100, Pb 300, Se 14, Zn 200. Other factors

266

taken into accounts when applying sewage sludge are the soil type, the type of culture,

267

the underground water level, and the distance from the nearby human settlement.

268

269

Currently sewage sludge are applied only over 10-20000 ha around the major cities

270

Pitesti, Suceava, and Timisoara (Borza and Radulescu, 2000). Badea et al. (2003)

271

estimate that 3% of the total sewage sludge production is used in agriculture, which is

272

16200 t for 2000 year.

273

274

In 1971 the composition of the domestic urban waste of Romania included 40-70 % as

275

vegetation and organic materials (Jianu and Alexandrescu, 1971). The percent decreased

276

since then down to 39% as food remnants in 2000 (Badea et al., 2003) but the potential

277

for preparing urban composts for agricultural use remained high. RDNIIEP (Research

278 and Development National Institute for Environmental Protection) is involved in an  
279 ongoing European project dealing with the development of a standard methodology for  
280 characterizing the composition of urban domestic wastes. Currently no composts are  
281 prepared on a large scale, and consequently urban composts are not used in the  
282 agriculture. However, one of the objectives of the strategy for wastes management is  
283 the reduction of biodegradable wastes sent to waste deposits, and activities of compost  
284 preparation and distribution to agriculture are explicitly stated (Badea et al., 2003).

285

286 In the 80<sup>s</sup> a number of experiments were done for preparing such compost and applying  
287 it in the field (Dumitru et al. 1991, 1993a, Vijiiala et al., 1994b). Data concerning  
288 metals concentrations are presented in table 5.

289

#### 290 1.2.4 Pesticides

291 3870 t of herbicides, 1240 t of insecticides and 432 t of fungicides have been applied in  
292 2000 over 4650110 ha (\*\*\*, 2001a). No data concerning the concentration of heavy  
293 metals in pesticides are available in the literature. Cu show the highest average  
294 concentration in the soil of vineyards (table 1), which might be due to the input related  
295 to applications of protective substances. However, these concentrations are rather low  
296 compared to other European countries, and not very much different from those found in  
297 other land uses types.

298

#### 299 1.2.5 Irrigation water

300 As mentioned in the introduction, the areas irrigated have strongly decreased in the last  
301 years. Data on heavy metals concentrations in the irrigation water are not available. An  
302 assessment of the input can be done using the concentrations stipulated for water of 2<sup>nd</sup>

303 class of quality, which is acceptable for irrigation, and an average water input of 3000  
304 m<sup>3</sup>/ha/year. It is worthy noting that the lack of irrigation infrastructure impinges also on  
305 the utilization of waste water as fertilizer, because, as we have seen, frequently a  
306 dilution is needed.

307

#### 308 1.2.6 Flood-related inputs

309 This type of input is highly relevant only for the Danube floodplain, especially in the  
310 context of the current reconstruction plans (Vadineanu et al., 2000). The patterns of  
311 metals distribution in Danube floodplain ecological systems have been recently  
312 analyzed (Iordache et al., 1998, Neagoe et al., 2000), as well as the fluxes associated  
313 with the flood and anthropic use (Iordache, 2002a). The net retention of metals in the  
314 floodplain landscapes was in the range of 6.4-22.9 kg/ha/year for Cu, 6.7-50 for Zn, 2.3-  
315 17.4 for Cr, 2.3-16.9 for Pb, and 0.053-0.168 for Cd, depending on the landscape type.  
316 This retention was mainly due to mechanical sediment filtration. However, a net export  
317 of the dissolved forms of Zn, Cr, and Cd was remarked, due to biological and  
318 physicochemical process involving the flooding water during the flood. Because of the  
319 ecotoxicological importance of dissolved metals, an assessment of the potential  
320 importance of this export after an eventual extensive restoration of the floodplain was  
321 done (Iordache, 2002b), leading to the conclusion that the restoration would not raise  
322 problems from this point of view more than the current natural systems, because the  
323 soils from the diked areas (used as agricultural land since the 60<sup>s</sup>) are not significantly  
324 different in terms of heavy metals, from the not diked soils of the Danube floodplain. At  
325 this stage, we will not take into consideration the inputs of heavy metals related to  
326 floods when assessing the balance, due to the low percent of agrosystems in the current  
327 Danube floodplain.

328

### 329 1.2.7 Livestock feeds

330 Only one source related to inputs of metals by livestock feed was found available in the  
331 literature (Gatlan et al., 1998). These source states that the concentrations in animal  
332 feeds analyzed in the period 1996-1998 (279 samples, as part of the sanitary control  
333 process) have not been higher than those acceptable at the that time (3 ppm for Cu, 50  
334 ppm for Zn, 0.5 ppm for Pb, and 0.1 ppm for Cd). Thus, the flux associated to animal  
335 feeds cannot be assessed. However, the input of heavy metals in livestock feeds is to  
336 some extent reflected in the animal manure input, which we have seen that can be  
337 estimated to some extent.

338

## 339 **2 Outputs**

### 340 2.1. Runoff and erosion

341 No data of runoff fluxes or superficial underground flows are published. Erosion by  
342 water and land sliding are responsible for an estimated average of 13.3 t/ha (range 3.2 -  
343 41.5 t/ha) soil losses (Dumitru et al., 2001). Quantification of erosion by wind is not  
344 available. Using this average flux and the surfaces affected by erosion by water and land  
345 sliding we would obtain a soil output of 126 milion t, which is 2-3 times higher than the  
346 current annual discharge of sediment into the Black Sea by the Danube river (Panin et  
347 al., 1996). Consequently, the average erosion rate of the 6300000 ha suffering this  
348 process might lower, or an important buffering of the fluxes occur transversally and  
349 longitudinally in the river basins, and we cannot estimate a realistic flux of metals by  
350 this mechanism.

351

### 352 2.2 Leaching

353 The available literature concerning leaching was limited to Barbu et al. (2004). The  
354 estimated fluxes (average values for 1998-2000 period) were the same as in the case of  
355 atmospheric deposition. Leaching of sulphur below 60cm depth ranged between 1  
356 kg/ha/year (at altitudes lower than 500m), and 7 kg/ha/year (at higher altitudes). At low  
357 altitudes sulphur had an accumulation trend in the soil, while at high altitudes the output  
358 was equal or even higher the input due to atmospheric depositions. The same pattern  
359 was noticed in the case of chloride and N-NO<sub>3</sub>, but not in the case of N-NH<sub>4</sub>, for which  
360 the leaching were always smaller than the atmospheric deposition. For total nitrogen the  
361 mentioned pattern applied as well, with less than 1kg/ha/year leaching at low altitudes.  
362 It is worthy noting that in 2000, a year with fewer than average precipitation, there was  
363 no leaching at all in the sites located at low altitudes (Barbu et al. 2004).

364

365 From the above information we can conclude, based on the reasonable assumption that  
366 the concentrations of heavy metals in the leaching water are smaller than those of S-  
367 SO<sub>4</sub>, that heavy metals fluxes are probably low at low altitudes (where the agrosystems  
368 are located) below 1 kg/ha/year. This is also indirectly pointed out by the fact that the  
369 analysis of underground water quality (MAPPM, 1996) shows contamination problems  
370 only in the case of N-NO<sub>3</sub>, to a smaller extent P-PO<sub>4</sub> and organic pollutants.

371

### 372 2.3 Biological products

373 Estimations of the fluxes associated with biological products are available in special  
374 cases related to experimental fields. Most of the already cited literature describing  
375 experiments with sludge, composts, manure, presents also the vegetal production  
376 enhancement and its quality in terms of metals concentration. However, no study has  
377 been published with regard to a real farm system. The main plant food products (\*\*\*,



378 2001) had 20508300 t in 2000 (this amount does not include the residual part of the  
379 plants, used in many cases for other purposes than human food), and the main animal  
380 food products had 1924300 t (residual products also not included). Using the acceptable  
381 concentrations for food products, one can compute the output of heavy metals by food  
382 products.

383

384 Plant and food products are regularly analyzed for heavy metal concentrations as part of  
385 the sanitary control (Dumitrescu et al., 1979, Popa and Stanescu, 1981). However, these  
386 data are not usually published, and when they are published the results are presented no  
387 actual concentration are shown, but only the situation related to the acceptable threshold  
388 values feeds (Gatlan et al., 1998, Vincu et al., 1998). For instance, from 2578 meat  
389 samples analyzed in two counties from 1996 to 1998, only two samples (beef kidney,  
390 1.5 ppm Cd) had concentrations higher than the acceptable values already mentioned in  
391 the subchapter concerning livestock feeds (Gatlan et al., 1998). The same authors cite  
392 Fantana (1998) as pointing out the need for geographically specific monitoring of food  
393 quality assisted by computer. From a different, nutritional, perspective, Vlad et al.  
394 (2000), publish a macro and micronutrients data set for the case of vegetal food in a  
395 limited geographic area of Romania. For computing the balance we will use the  
396 acceptable limits of heavy metals in animal food in 2000 (0.1 ppm Cd, 0.5 ppm Pb, 3  
397 ppm Cu, and 50 ppm Zn, based on The Health Ministry Order 611/1995, no longer in  
398 use in 2003), and the acceptable limits in plants mentioned by Vrinceanu et al., 1997 (15  
399 ppm Cu, 100 ppm Zn, 10 ppm Pb, 1 ppm Cd). This will be an overestimation of the  
400 output fluxes mediated by food products in the case of these metals, if we consider that  
401 the food products had concentrations below the acceptable limits. On the other hand,  
402 neglecting the residual products will induce an underestimation of the overall fluxes

403 mediated by biological compartments. Concentration of Mn, and Cr in plant food  
404 products will be estimated according to Vlad et al. (2000), and that of Co according to  
405 \*\*\*, 1981-87.

406

#### 407 2.4 Animal excreta and waste water

408 According to our estimations 28359882 t of animal excreta are produce, while just  
409 15812625 t are applied as natural fertilizers. As long as it is stored outside of the  
410 agricultural area, the difference can be considered as an output. But it is more likely to  
411 be stored in the nearby of the animal farms. They might be responsible for the 1149 ha  
412 excessively polluted as a result of animal wastes application (table 2). If we consider  
413 that formally this area is inside the agricultural land, we can not consider the 12547257 t  
414 of animal wastes as an output. However, if we consider that this is in fact a flow with  
415 origin in the 14329900 ha of not permanent cultures which sustain the animal raising  
416 (with the underlying assumption that the animal feed metal import at national scale is  
417 not higher than the export), and going to the 1149 ha of excessively polluted sites, we  
418 can consider this flow at least as a very important internal flow at national scale. At this  
419 stage of the national balance we will consider, for simplification, and taking into  
420 consideration the very small surface of the excessively polluted sites (0.006% of the  
421 total) that the flux of animal wastes not used as fertilizer is actually an outflow from the  
422 total agricultural area.

423

424 There is also a production of waste water, estimated at 88790426 m<sup>3</sup>, which we have  
425 seen that cannot be applied in the field as fertilizer without dilution in the frame of a  
426 functional irrigation system. Even if this water would be dumped on the excessively  
427 polluted sites, it is very likely that it will leave in some way the land towards aquatic

428 systems or wetlands, either by leaching or by runoff. More probably this water is  
429 released in wetlands or even aquatic system (either directly or by a sewage system).  
430 Consequently we will consider the animal waste water as an output. It will be a partial  
431 double counting here in terms of amounts, because part of the quantity of waste water is  
432 due to urine, which has been already included in the animal wastes.

433

### 434 **3 Heavy metal balance for Romania**

435 Figure 1 presents a highly simplified homomorphic model used for computing the  
436 balance for the agricultural lands of Romania. Table 8 includes the values of the fluxes  
437 and the resulting net balance.

438 *Figure 1 about here*

439 The assumptions underlying the evaluation of the fluxes have been already presented in  
440 the case of each analyzed media. According to the resulted balance, the major inputs of  
441 metals to agricultural lands are by natural fertilizers and atmospheric depositions, and  
442 the major outputs occur by dumped animal wastes and main plant food products.

443 Several aspects which would presumably influence the above conclusions have been  
444 completely neglected, namely:

- 445 • plant food products does not include the residual part of plants, which in many cases  
446 is used as fodder in animal farms. Thus, the export by plant products is much  
447 underestimated. Normally this export should be higher than the input by natural  
448 fertilizers, if the metals in animal feed has their main origin in the local plant  
449 products.
- 450 • how much of the output by dumped animal wastes and main animal food products is  
451 originating in animal feed (from industry and trading) with foreign origin and in  
452 contamination within the animal farm. We have assumed that all this flux has origin

453 in local agricultural land (by plant food products transferred to animal farms by  
454 industry and trading). If the foreign animal feed is important, the outputs by wastes  
455 are over estimated.

- 456 • how much of the input by natural fertilizers has different origin than animal feed  
457 produced by using local plant products.
- 458 • how much export by erosion takes place. This is probably a major export pathway in  
459 Romania. From this point of view the output is strongly underestimated
- 460 • what is the real value of atmospheric deposition of metals in Romania.

461

462 Because the surfaces on which some inputs apply are in fact much smaller than the total  
463 surface of the agricultural land, and because of the high variability in space of the  
464 atmospheric depositions, one can expect very large differences from this balance, at  
465 regional and local level.

466

*Table 8 about here*

#### 467 **4 Research directions**

468 One can identify two kinds of research priorities:

- 469 • one is related to strictly improving the possibility for computing a metal balance for  
470 the agricultural land at different level (from farm to national level),
- 471 • and the second one is related to putting the problem of heavy metals pollution of the  
472 agricultural land in the broader context of heavy metals biogeochemistry and  
473 ecotoxicology at landscape level, and of the diversity of problems related to  
474 agricultural land, and more generally, to land use.

475

476 Table 9 present the first kind of research priorities. In what concerns the second kind of  
477 priorities, in the introduction, where the context of the analyzed problem was described,

478 we have pointed out the main limitations in heavy metals research in Romania, as linked  
479 to the dominant sectoral approach. It has been also shown that heavy metals pollution is  
480 only one of problems relevant for agrosystems. Thus, if one would be to identify the  
481 main research priorities from this integrating perspective, these would be:

- 482 1. research programs in landscape biogeochemistry, nutrition science and  
483 ecotoxicology.
- 484 2. integrated approach of land use problems in order to assess the sustainable regional  
485 development

486

487 Figure 2 presents (after Iordache 2002a) the possible co-ordinates of a research program  
488 relevant for the first direction, which would be extremely useful to be developed in the  
489 highly polluted landscapes of Romania, by a consortium of national partners.

490

491 To follow the second direction one would have to develop tools and method as part of  
492 the decision support system for the codevelopment of the natural capital and socio-  
493 economic systems (Vadineanu, 1999), tools able to surpass the current sectoral decision  
494 making and decision support. It has been advocated such an approach for instance in the  
495 case of the national monitoring system (Iordache, 2002c, d) or for the risk assessment  
496 regulation (Iordache, 2002e).

497

*Table 8 about here*

498

*Figure 2 about here*

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503

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767        [/.%5C20\\_Romania.pdf](#)
- 768



769 **Figures legends**

770 **Figure 1** Simplified homomorphic model used for computing the balance for the  
771 agricultural lands of Romania. Legend: 1 urban sewage sludge, 2 natural fertilizers, 3 N,  
772 P, K fertilizers, 4 atmospheric depositions, 5 irrigation, 6 animal wastes not used as  
773 fertilizers, 7 animal waste water, 8 main plant food products, 9 leaching, NE not  
774 evaluated.

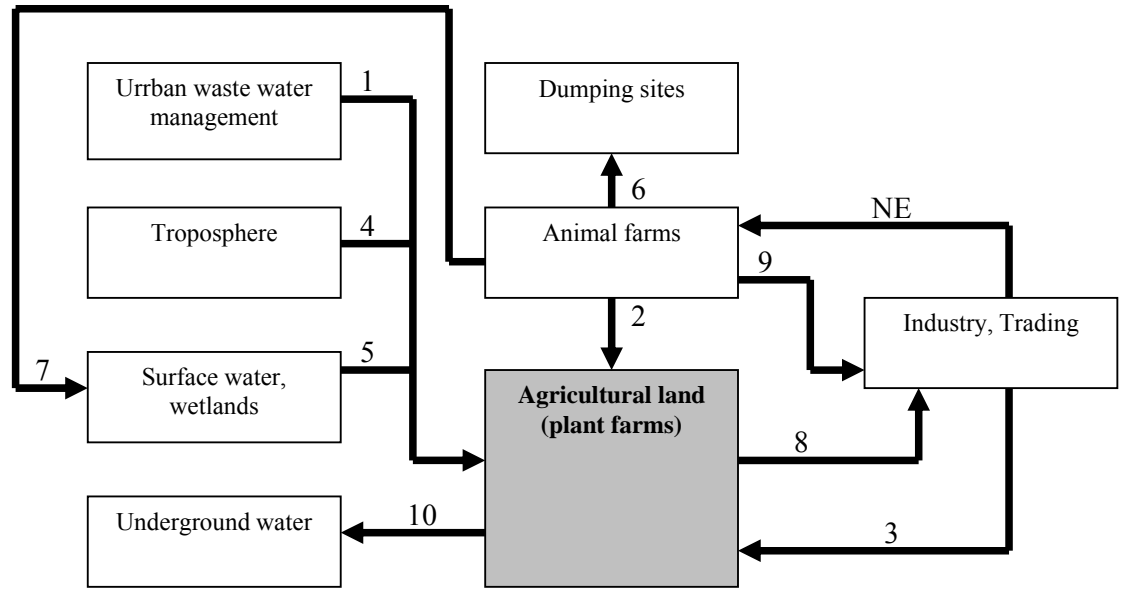
775

776 **Figure 2** Conceptual framework for a research program in landscape biogeochemistry,  
777 nutrition science and ecotoxicology. *Legend:* Black arrows = fluxes of elements, gray  
778 arrows = control pathways, squares = systems, circles = control parameters.

779

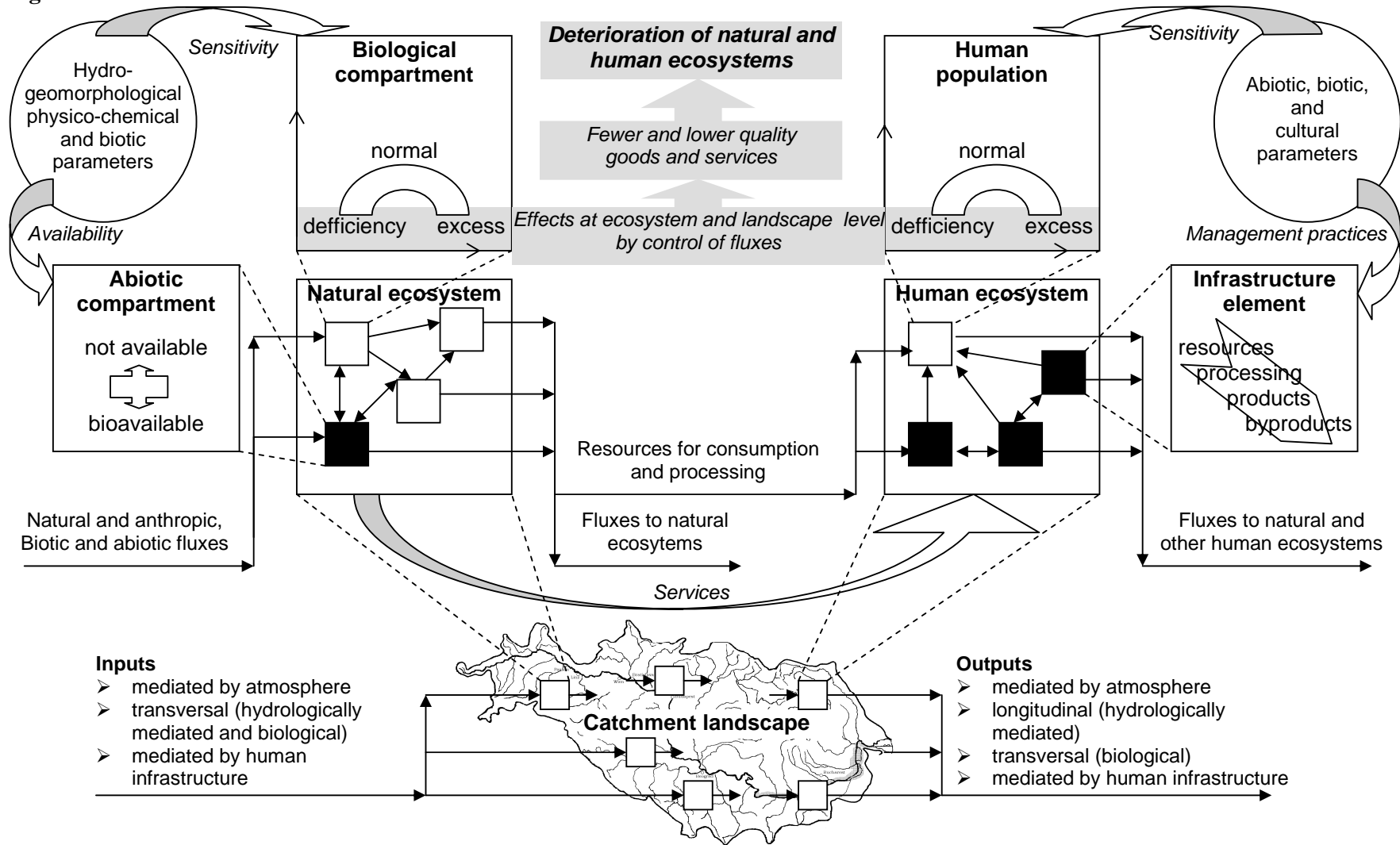
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**Figure 1**



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**Figure 2**



803 **Table's caption**

804 **Table 1** Surfaces with productive capacity affected by limiting factors. In bold are typed  
805 the factors related to metals (adapted after Dumitru et al., 2001).

806 *Table 1*

807 Legend:

808 <sup>1</sup> One and the same surface can be affected by one ore more limiting factors

809

810 **Table 2** Areas covered by excessively polluted agricultural land in 1992 (after  
811 Vrinceanu et al., 1997). In bold, sources related to agriculture.

812

813 **Table 3** Contents (ppm d.w.) of potential polluting elements in top soil of monitoring  
814 sample plots by main agricultural land uses (A; surfaces in 2000 given in thousands of  
815 ha, n = number of sampling stations) and soil types (B). After Dumitru et al. (2000).

816

817 **Table 4** Assessment of total animal wastes produced in 2000.

818

819 **Table 5** Chemical parameters of animal manure, sludge and composts.

820

821 **Table 6** Chemical parameters of industrial and urban sludge and composts (\* = average  
822 for 3-5 consecutive years).

823

824 **Table 7** Chemical parameters of animal farm waste water

825

826 **Table 8** Estimated fluxes and balance for agricultural lands of Romania (g/ha  
827 agricultural area in year 2000), the land surface relevant for the considered fluxes, and a

828 short characterization of the data used for computation.

829

830 **Table 9** Priorities for future research activities in order to improve the knowledge

831 concerning heavy metal balance in agricultural ecological systems.

832 *Table 9*

833 Legend:

834 <sup>a</sup> available knowledge: 1 = none, 2 = little, 3 = sufficient

835 <sup>b</sup> Kind of research/information needed: e.g. fundamental data, harmonisation,  
836 demonstration in practice, etc.

837 <sup>c</sup> relevance for balance outcome: 1 = low/not relevant, 2 = medium, 3 = high; ? =  
838 difficult to assess from the country's point of view

839 <sup>d</sup> research priority: 1 = none, 2 = little, 3 = medium, 4 = high, 5 = very high priority

840

841 **Table 1**

Factor	Affected surface <sup>1</sup> (thousands of ha)	
	total	of which arable land
Frequent drought	7100	
Periodic soil humidity excess	3781	
Soil erosion by water	6300	2100
Soil erosion by wind	378	
Landsliding	702	
Soil salinization	614	
Soil compaction	6500	6500
Topsoil hard layer	2300	2300
Low and very low humus pool	7485	4525
Moderate and high acidity	3424	1867
Low and very low burden of mobile P	6330	3401
Low and very low burden of mobile K	787	312
Low and burden of N	5110	3061
<b>Deficiency in microelements (Zn)</b>	<b>1500</b>	<b>1500</b>
<b>Chemical pollution of soil, of which</b>	<b>900</b>	
<b>excessively polluted</b>	<b>200</b>	
polluted with oil and salted water	50	
<b>polluted by air dispersal</b>	<b>147</b>	
Soil destruction by excavations	15	
Soil covering by solid wastes	18	

842

843

844 **Table 2**

Pollution causes	Area (ha, %)		Polluting economical branch
Emissions caused by compounds of heavy metals, nitrogen and sulfur oxides, dusts and gases	53179	31.422	chemical and petrochemical industry
Releases of oil, salty water and derrick sludge	49368	29.170	oil extraction industry
Emissions of gases and deposits of dusts with metal compounds	28251	16.693	metallurgical industry
Emissions in air of some calcareous dusts resulted from plants building materials and refractory materials	15516	9.168	building material industry
Emissions of sulphur and nitrogen oxides, release of ashes from spoil banks taken by wind	11806	6.976	thermo-energetic industry
Deficiencies in the activity of ore extraction processing	7453	4.404	mining industry
Industrial dusts, wastes and waste water	300	0.177	mechanical engineering industry
Dusts of ferric oxides	300	0.177	
Storage of industrial wastes	15	0.009	
Storage of ballast products	14	0.008	
<b>Release and deposition of animal wastes from zootechnical farms</b>	<b>1149</b>	<b>0.679</b>	<b>agriculture</b>
Waste waters, dirt and wastes	11	0.006	food industry
Storage of dirt, wastes and industry	1880	1.111	local administration

845

846 **Table 3**847 *A (range and average)*

Element	Arable (9381.1 n=437)		Vineyards (272.3 n=18)		Orchards (254.6 n=16)		Pastures (3441.7 n=140)		Meadows (1507.1 n=59)	
	range	average	range	average	range	average	range	average	range	average
Cu	8-220	24	10-125	34	15-88	29	5-27	21	5-80	26
Pb	7-92	28	10-45	29	10-45	29	6-170	32	12-95	30
Zn	23-500	87	45-250	98	48-460	139	10-360	88	33-475	122
Cd	0 (udl)-2.5	0.9	0 (udl)-1.5	0.7	0.4-2	1	0 (udl)-5	1	0 (udl)-3	1.1
Co	3-35	18	5-25	17	6-35	18	2-25	17	5-40	19
Ni	7-85	34	15-60	33	13-55	37	5-75	32	10-105	35
Mn	120-1195	505	290-890	496	310-670	473	30-1193	458	125-1326	481
Cr	10-193	54	15-98	52	30-103	52	5-154	49	14-430	54
S soluble	22-1036	209	77-393	201	121-1036	301	14-2046	254	30-460	189

848 *B (average)*

	vertisols	molisols	argiluvissols	cambisols	undeveloped soils	spodosols	umbrisols	hydromorphic soils	halomorphic soils
Cu	26	24	22	24	25	19	20	25	20
Pb	29	27	29	32	31	35	25	34	11
Zn	127	96	82	101	97	57	157	101	76
Cd	1	0.9	0.8	1.1	0.9	0.8	1.7	1	1.8
Co	18.7	18	16	19	19	12	25	17	5.5
Ni	33	35	27	37	37	25	32	37	21
Mn	469	504	497	486	492	270	455	501	373
Cr	51	56	47	51	56	39	45	57	46
S soluble	245	206	204		265	247	131	231	381

849



850 **Table 4**

		No of animals in 2000 (thousands)	Waste water* l/animal/day	waste (urine and feces)		Total waste (t/year)		Total waste water (m3/year)**
				t wet/animal/year	t dry/animal/year	wet mass	dry mass	
Cattle	Cows	1769	30	14.94	1.89	2642886	334341	19370550
	Other	1282	30	6.7	0.77	858940	98714	14037900
Subtotal cattle		<b>3051</b>				<b>3501826</b>	<b>433055</b>	<b>33408450</b>
Pigs		5848	25	2.38	0.21	1391824	122808	53363000
Sheeps		7906	4	0.73	0.18	577138	142308	***
Goats		558						
Horses		858						
Poultry	For eggs	38497	0.08	3.86	0.96	14859842	3695712	1124112
	Other	30646	0.08	2.62	0.65	8029252	1991990	894863
Subtotal poultry		<b>69143</b>				<b>22889094</b>	<b>5687702</b>	<b>2018976</b>
<b>GENERAL TOTAL</b>						<b>28359882</b>	<b>6385873</b>	<b>88790426</b>
Source		***, 2001a	Jinga et al, 1990	Rauta and Dumitru, 1986	Rauta and Dumitru, 1986			

\* technological changes may have changed the amounts currently produced

\*\* where computed, we assumed that 100% of the animals are grew in animal farms, which is over estimation

\*\*\* not computed because percent of sheeps grew in industrial animal farms is presumably very far from 100%, due to traditional raising in the mountains

853 **Table 5**

parameter	dimension	Solid manure					Animal sewage sludge					Animal sewage sludge compost		
		sheep	poultry	poultry	cattle	cattle	pig	pig	pig	pig	poultry	pig	pig	pig
water content	% fm		39.8					66.8		82	69.01			
pH	-		5.3						7	6.23	8.04			
Cd-content	mg/kg dm	0.3	5.35	6.3	5.2	1.67		0.04	0.055		3.05	4.4	200	2.9
Cr-content	mg/kg dm	6.4	84.5	65	190			6.1	42.6		229.5	175	1133	54
Cu-content	mg/kg dm	7.7	254.5	100	40	20	7	39	81	7	97.5	138	602	213
Ni-content	mg/kg dm	4.3	7	3.8	2.7	14					14.65	4.7	117	3.7
Pb-content	mg/kg dm	4	90	30	28	29		5.7	5.7		69.5	51		20
Zn-content	mg/kg dm	230	414	390	250	101	42	113	101.35	42	1487	454	2103	497
N-total	% dm	0.98	0.92	2.15	1.44	1.84	3.4	2.42	3.015	3.4	0.755	0.88	1.72	1.67
P-total	% dm	0.57	0.745	1.7	0.7	0.52	1.09		0.57	1.09	1.48	0.61	1.43	
K-total	% dm	0.9	0.545	1.9	2.68		0.23		0.155	0.23	3.34	0.88	0.67	
Mg-total	% dm	0.72	0.64	0.69	0.58			0.29			1.455	0.3		0.41
Ca-total	% dm	0.72	1.83	0.77	0.57			1.17	1.065		3.35	0.65	1.03	1.39
C-total	% dm							30.6	25.72					
NO <sub>3</sub> -N-(CaCl <sub>2</sub> )-soluble	mg/100g dm	321		11	120.6			22.5	103	194	1.9	11.6		442
NH <sub>4</sub> -N-(CaCl <sub>2</sub> )-soluble	mg/100g dm	152	45.55	1.5	0.5			33.2		962	7.875	0.1		293
P-(lactate)-soluble	mg/100g dm													
K-(lactate)-soluble	mg/100g dm	2587	905	1000	660			3			260.5	42	5.75	
Mg-(CaCl <sub>2</sub> )-soluble	mg/100g dm	185	176.5	18.7	28.9			51			102	28.9		
soluble SO <sub>4</sub> -S	mg/100g dm	3136	6223.5		946.4			113			1475	210.9		
B	mg/kg dm													13
Mn	mg/kg dm	681	1270	740	430	267		312			492.5	822	390	757
Fe	mg/kg dm	1396	9200	2030	6290				524.5		13359	16900	24460	17960
Na	mg/kg dm		2650					672	1500	780	11000			
Cl	mg/kg dm	1335		1636	1087			125			968.5	81		
Co	mg/kg dm	1.9	19	8.8	8.2	8		2.7	0.73		6.575	14		29
Ca <sup>++</sup>	mg/kg dm	190	950	900	210			510			205	688	159	
Data source		Dumitru et al., 1990	Dumitru et al., 1989c	Dumitru et al., 1989c	Dumitru et al., 1989c	***, 1981-87	Dumitru, 1983	Dumitru et al., 1993c	Dumitru et al., 1986	Nastea et al., 1986a	Dumitru et al., 1989c	Dumitru et al., 1989c	***, 1981-87	***, 1988, nr 7

Table 6

parameter	dimension	Industrial sewage sludge				Urban sewage sludge					Urban sewage sludge compost		Urban household wastes compost		
		celulose industry	celulose industry	leather industry	chemical industry	Bucharest	Bucharest*	Cluj-Napoca	Bucharest	Timisoara*	Bucharest	Bucharest*	Bucharest*	Bucharest	Bucharest
water content	% fm				75.6		53.75		57.43	49.3		38.85			
pH	-				8.42		7.5		7.17	5.8			8.28	8.3	8.45
Cd-content	mg/kg dm			13	4	68	38	32.5	48	23	200	23.9	12.5	13	12
Cr-content	mg/kg dm			3916	9	3485	505	2567	760	982	1135	833.75	189	224	150
Cu-content	mg/kg dm	139.93	224	310	242	841	473	352	558	509	602	3919.5	168	129	173
Ni-content	mg/kg dm			5.1	22	18.8	102	103	21		117	69	45	47	35
Pb-content	mg/kg dm	114	217.5	202	119	300	238	801	375	261	410	231	394	365	320
Zn-content	mg/kg dm	286.67	227.33	1692	497	3910	2025	1145	3240	619	2103	1894	1285	250	1780
N-total	% dm	2.18	2.76	0.98	8.23	0.51	0.72	3	0.53	0.97	1.72	0.615	1.57	1.67	1.35
P-total	% dm	0.99	0.93	0.116	1.5	0.82	0.66	1.1	1.16	0.58	1.43	2.155	0.29	0.36	0.21
K-total	% dm	1.5	1.21	0.035	1.8	0.96	0.91	0.7	1.4	1.42	0.07	5	1.03	1.5	0.53
Mg-total	% dm			0.071	2.01	0.48	0.77		0.97	0.3	0.43	1.375	0.28	0.39	0.16
Ca-total	% dm	0.2	0.48	1.9	15.2		0.66		0.65	0.35	1.03		2.18	2.95	1.35
C-total	% dm						11.56		24.53	36		47.35		37.4	
NO3-N-(CaCl2)-soluble	mg/100g dm			35,7		0.875	8		9		1.65	43.2	8.9	12.8	5.2
NH4-N-(CaCl2)-soluble	mg/100g dm					2.431	14.7		22.2		11.4	27.585	31.3	0.1	71.9
P-(lactate)-soluble	mg/100g dm	5.464	2.804				22.3					18.4			
K-(lactate)-soluble	mg/100g dm	7.576	40.716	2.6	5.25	15.5	44		5.32		5.75	9.36	351.2	440	281
Mg-(CaCl2)-soluble	mg/100g dm			9.93		16.5	9.3		1.3		0.015	7.45	209	23	18.7
soluble SO4-S	mg/100g dm	1311.3	1964.83	769.7		936.1	2062.3		4492.2			2415.85	748.5	845.2	677
Mn	mg/kg dm	51.67	203.83	258	212	605	492	536	630	214	350	468.9	523	400	680
Fe	mg/kg dm	4233.33	5586.33	25818	5520	25155	24843	20790	24167		24460	22638.5			
Na	mg/kg dm				9700		3600		2100			14150	8000	12500	5300
Cl	mg/kg dm			856.1		80.65						1790.15	5370	6581	4309
Co	mg/kg dm	2.78	12.16	8.9	9	31.5	24	29.01	28	25	16	24.5	12	12	11
Ca++	mg/kg dm			1197		125			1260		159	2300	1288	1500	1025
Data source		***, 1981-87	***, 1981-87	***, 1981-87	Dumitru et al., 1984	***, 1981-87	Dumitru et al., 1993b	Blaga et al., 1989	Dumitru et al., 1994	Borza and Radulescu, 2000	***, 1981-87	Vijiiala et al., 1994a	Vijiiala et al., 1994b	Dumitru et al., 1991	Dumitru et al., 1993a

858 **Table 7**

parameter	dimension	cattle	pig	pig	pig	pig	pig	poultry	poultry
pH	-		8.42	7.5	7.78				
Cd-content	mg/l	0.07	0.016		0.02		0.027	0.01	0.11
Cr-content	mg/l	0.06	0.072		0.059		0.067	0.16	0.24
Cu-content	mg/l	0.5	0.01		0.04		0.1	0.15	0.08
Ni-content	mg/l	1.1	0.037		0.018		0.26	0.015	0.014
Pb-content	mg/l	0.4	0.087		0.239		0.434	0.09	0.36
Zn-content	mg/l	1.1	0.48		0.74		0.96	1.34	1.26
N-total	mg/l	468	250	430	323	430	521	196	207
P-total	mg/l	226	36.08	71	47	60	56	8.82	26.6
K-total	mg/l	1625	201	231	224	220	220	56	109
Mg-total	mg/l	1623	38.9		50		71	43.17	0.04
Ca-total	mg/l	596	89		104		108	106	110
NO3-N	mg/l			0.56	3.72		6.2	2	11.5
NH4-N	mg/l	320	52.45	178.2	567		618.3	23.54	10
K+	mg/l	28.6		120	216.45	129.5	242.58	60.5	90
Mg++	mg/l	6.4		51	51.96	59	67.92	42	33
SO4-S	mg/l	3.6		580	94.8	550	114.4	308.17	0
Mn	mg/l	1.2	0.025		0.12		0.028	0.76	0.63
Fe	mg/l	1.4	4.3		5.51		5.74	8.31	7.07
Na	mg/l		105	280	104				
Cl-	mg/l	23.4		425	182.87		219	198.67	663
Co	mg/l	0.2	0.055		0.049		0.044	0.003	0.05
Ca++	mg/l	11.2		29	68	30	70.4	77.83	125
Data source		Dumitru et al., 1989d	Nastea et al., 1986b	Nastea et al., 1986a	Dumitru, 1991	Rauta et al., 1980	Dumitru et al., 1989a	Dumitru et al., 1989c	Dumitru et al., 1989c

859

860 **Table 8**

<b>Inputs</b> (as over 14856800 ha)	Cu	Zn	Co	Ni	Mn	Cd	Pb	Cr	Real surface of input / output (ha)	Source of data: concentrations, amount or fluxes
Urban sewage sludge	0.596	2.386	0.030	0.067	0.540	0.046	0.431	1.810	15000	conc and amount (average Ro literature)
Natural fertilizers	69.576	337.432	7.438	8.240	637.461	2.922	34.844	94.893	647200	conc and amount (average Ro literature)
N fertilizers	0.053	0.157	NE	0.360	0.032	0.064	0.122	0.045	3724578	conc (average European literature), amount (Ro literature)
P fertilizers	0.134	1.450	0.012	0.321	0.594	0.099	0.122	0.933	3724578	conc (average European literature), amount (Ro literature)
K fertilizers	0.002	0.006	NE	0.016	0.018	0.003	0.010	0.002	3724578	conc (average European literature), amount (Ro literature)
Atmospheric deposition	2	100	1.4	7.5	17.5	2.5	40	NE	14856800	fluxes (half of the average for Hungary)
Irrigation	0.825	4.123	NE	2.062	NE	0.041	0.206	2.062	204200	conc (Ro regulations, water class 2), amount (as described in the report)
<b>Total input</b>	<b>73.185</b>	<b>445.553</b>	<b>8.880</b>	<b>18.565</b>	<b>656.145</b>	<b>5.675</b>	<b>75.734</b>	<b>99.744</b>		
<b>Outputs</b> (as from 14856800 ha)										
Animal wastes	55.208	267.751	5.902	6.538	505.823	2.319	27.648	75.297	14329900	conc (Ro literature) amount (estimation based on Ro literature)
Animal waste water	0.877	5.857	0.399	1.438	2.752	0.252	1.604	0.655	14329900	conc (Ro literature) amount (estimation based on Ro literature)
Main plant food products	20.706	138.040	1.380	1.063	6.902	1.380	13.804	0.690	14856800	conc and amount (regulations and Ro literature)
Main animal food products	0.389	6.476	NE	NE	NE	0.013	0.065	NE	14329900	conc and amount (regulations and Ro literature)
Leaching	<1	<1	<1	<1	<1	<1	<1	<1	NE	fluxes (Ro literature as described in the report)
<b>Total output</b>	<b>77.179</b>	<b>418.124</b>	<b>7.682</b>	<b>9.039</b>	<b>515.477</b>	<b>3.964</b>	<b>43.121</b>	<b>76.643</b>		
<b>Balance</b>	<b>-3.994</b>	<b>27.430</b>	<b>1.198</b>	<b>9.526</b>	<b>140.668</b>	<b>1.711</b>	<b>32.613</b>	<b>23.101</b>		

862 **Table 9**

Research field	Knowledge/ data available (1 - 3) <sup>a</sup>	What is needed? -- <sup>b</sup>	Relevance for balance (1 - 3; ?) <sup>c</sup>	Research priority (1 - 5) <sup>d</sup>	comments
<b>Fundamental research</b>					
analysis, methodology	2	improvement and harmonisation of existing knowledge	2	3	focused by now on N and P balances in fluvial systems and in agricultural landscapes
soil research	3	GIS development	2	3	running research initiatives
<b>Heavy metal inputs</b>					
material research	2	fundamental data	?	2	
animal nutrition	2	fundamental data	1	4	by now only in terms of needed micronutrients, not in terms of pollution with heavy metals
animal medication	1	fundamental data	1	2	
mineral fertilisers	2	fundamental data	3	3	
atmospheric deposition in rural areas	2	improvement of the monitoring	2	5	by now only related to the pan-European monitoring of forest ecosystems
special crops	3	GIS development	3	4	
inputs special for organic farming	2	fundamental data	2	2	before 1989 pilot studies in experimental fields
<b>Heavy metal outputs</b>					
leaching	1	fundamental data	2	4	
erosion/surface run off	2	fundamental data	3	5	known only the surfaces affected by erosion in different degrees, but no quantification of fluxes
soil-plant-transfer/food chain	3	GIS organization	3	3	running research initiatives
removal of soil with crops	1	fundamental data	2	2	
<b>Balancing</b>					
mass flow determination	2	fundamental data	3	5	limited to experimental fields
models for unknown fluxes	1	models development	3	4	most needed would be GIS models of atmospheric deposition and erosion
balancing approaches	2	studies for metals in agricultural landscapes	3	5	done for metals in the Danube floodplain, and for N & P in the Danube floodplain and in agricultural catchments
<b>Monitoring</b>					
organic farming	2	design of monitoring	2	3	organic farming still in development
farm level monitoring	1	design of monitoring	3	5	
livestock feeds	1	design of monitoring	3	4	
animal manure, industrial and bio-waste	2	improvement of monitoring	3	4	existing in terms of general quantities, but lacking in terms of subtypes and contamination with metals